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Author contributions

Arnemo: Conceptualization; **Fuchs, Thiel:** Methodology; **Rodushkin:** Resources; **Arnemo, Boesen, Evans, Fuchs, Græsli, Hydeskov, Rodushkin, Thiel:** Investigation; **Fuchs, Thiel:** Formal analysis; **Fuchs:** Writing – Original Draft; **Arnemo, Brown, Evans, Fuchs, Græsli, Hydeskov, Kindberg, Rodushkin, Thiel, Zedrosser:** Writing – Review & Editing; **Arnemo, Zedrosser:** Supervision; **Arnemo, Kindberg:** Project administration; **Arnemo, Kindberg:** Funding acquisition

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1 **High concentrations of lead (Pb) in blood and milk of free-ranging brown bears (*Ursus***
2 ***arctos*) in Scandinavia**

3

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24 **Abstract**

25 Exposure to lead (Pb) is a global health problem for both humans and wildlife. Despite a dramatic
26 decline in human Pb exposure following restrictions of leaded gasoline and industry and thereby an
27 overall reduction of Pb entering the environment, Pb exposure continues to be a problem for wildlife
28 species. Literature on scavenging terrestrial mammals, including interactions between Pb exposure
29 and life history, is however limited.

30 We quantified Pb concentration in 153 blood samples from 110 free-ranging Scandinavian brown
31 bears (*Ursus arctos*), 1-25 years old, using inductively coupled plasma sector field mass spectrometry.
32 We used generalized linear models to test effects of age, body mass, reproduction status and spatial
33 distribution on the blood Pb concentrations of 56 female bears. We sampled 28 females together
34 with 56 dependent cubs and paired their blood Pb concentrations. From 20 lactating females, we
35 measured the Pb concentration in milk.

36 The mean blood Pb concentration was 96.6 µg/L (range: 38.7-220.5 µg/L). Both the mean and
37 range are well above established threshold concentrations for developmental neurotoxicity (12
38 µg/L), increased systolic blood pressure (36 µg/L) and prevalence of kidney disease in humans (15
39 µg/L). Lactating females had higher Pb blood concentrations compared to younger, non-lactating
40 females. Blood Pb concentrations of dependent cubs were correlated with their mother's blood Pb
41 concentration, which in turn was correlated with the Pb concentration in the milk.

42 Life-long Pb exposure in Scandinavian brown bears may have adverse effects both on individual and
43 population levels. The high blood Pb concentrations found in brown bears contrast the general
44 reduction in environmental Pb contamination over the past decades in Scandinavia and more
45 research is needed to identify the sources and pathways of Pb exposure in the brown bears.

46 **Keywords**

47 Lead, Blood, Milk, *Ursus arctos*, Carnivora

48

49 Introduction

50 Lead (Pb) is a highly toxic element without any known biological functions, that negatively
51 affects multiple physiological systems in vertebrates (Bellinger et al., 2013). In a risk assessment, the
52 European Food Safety Authority (EFSA) established benchmark dose levels of blood Pb
53 concentrations for developmental neurotoxicity in children (12 µg/L) as well as for increased systolic
54 blood pressure (36 µg/L) and prevalence of kidney disease (15 µg/L) in adults. The EFSA further states
55 that there is “no evidence for a threshold of lead-induced effects” and defines tolerable intake levels
56 as not appropriate (EFSA, 2013).

57 Pb exposure in terrestrial mammals results from ingestion or inhalation. In humans ingested
58 or inhaled Pb initially increases blood Pb concentration with a relatively short half-life of
59 approximately 35 days (Rabinowitz et al., 1976). The organism treats Pb as a substitute for calcium
60 and transfers it to soft tissues and bones (Rabinowitz et al., 1976). In mammals, over 90% of the total
61 body Pb is stored in bones and teeth with a half-life of 10-30 years (Andreani et al., 2019; Rabinowitz
62 et al., 1976). During periods of nutritional stress, such as pregnancy and lactation, Pb stored in bones
63 and soft tissue can become an endogenous source resulting in increased blood Pb concentrations
64 (Silbergeld, 1991). Increased calcium demands during skeletal development of the fetus as well as
65 during lactation lead to increased calcium turnover during pregnancy and lactation in the mother. As
66 a result, Pb is released from bones into the blood and into the milk (Ettinger et al., 2014), which may
67 have detrimental effects on both the mother and her offspring.

68 In Europe, Pb has been mined, refined and used for more than 2,000 years, resulting in
69 widespread airborne pollution from smelting processes and, since the mid-20th century, from Pb as
70 gasoline additive (Settle and Patterson, 1980). In northern European lake sediments, the current
71 environmental Pb concentrations are up to 1000 times higher than the natural background
72 concentrations (Renberg et al., 2001; Settle and Patterson, 1980). Top soils are contaminated with Pb
73 globally, and the Pb uptake in plants growing in these soils pose a risk for consumers (Khalid et al.,
74 2017). Global environmental and health concerns have led to a gradual phasing out of leaded
75 gasoline since the 1970s, which accelerated in the mid-1980s, when European Union member states
76 started to reduce the allowed Pb limits in gasoline (von Storch et al., 2003). Consequently,
77 atmospheric deposition of Pb decreased, especially in northern Europe (Lind et al., 2006; von Storch
78 et al., 2003). For example, Danielsson and Karlsson (2015) reported that the mean Pb concentration
79 in mosses had decreased by 96% in Sweden from 1975 to 2015. Blood Pb concentrations of children
80 in southern Sweden decreased from 60 µg/L in 1978 to 11 µg/L in 2007 (Skerfving et al., 2015;
81 Strömberg et al., 2008) and to 8.5 µg/L in 2019 (data available at Karolinska Institutet, 2020). Liver Pb

82 concentration of bank voles (*Myodes glareolus*) in central Sweden measured in 2017 had decreased
83 by two thirds compared to 2001 (Ecke et al., 2020). This general decrease in exposure suggests a
84 direct link between aerial Pb pollution and Pb exposure in humans and the environment.

85 In high-income countries, most sources of Pb emission are strictly regulated today. An
86 exemption is Pb used in hunting ammunition, which presents a significant source of exposure for
87 both humans and wildlife (Arnemo et al., 2016; Bellinger et al., 2013). For example, Pb from hunting
88 ammunition is an important source of morbidity and mortality in avian scavengers, such as golden
89 eagles (*Aquila chrysaetos*) (Ecke et al., 2017) and white-tailed eagles (*Haliaeetus albicilla*) (Helander
90 et al., 2009). Increased Pb concentrations in birds also result in behavioral alterations, lower
91 reproductive success, and physiological changes (Berglund et al., 2010; Ecke et al., 2017; Finkelstein
92 et al., 2012; Kelly and Kelly, 2005). Generally, the extent of Pb exposure depends on a species'
93 feeding ecology and can vary within a population, depending on individual spatio-temporal
94 movement patterns and resource specialization (Arrondo et al., 2020; Brown et al., 2019;
95 Nadjafzadeh et al., 2013). Periods of nutritional stress, such as pregnancy or incubation, may
96 mobilize Pb from endogenous sources and increase the risk for clinical effects of Pb concentration in
97 wildlife (Lam et al., 2020).

98 Scientific evaluations of Pb exposure in wild-living terrestrial mammals commonly use a
99 screening approach, typically sampling of soft tissues, with human food safety or biomonitoring as
100 the main motivation (eg. Chiari et al., 2015; Morales et al., 2011). Studies including other variables of
101 the investigated populations, such as sex and age, vary in results. For example, female European roe
102 deer (*Capreolus capreolus*) in Spain had higher Pb concentrations than males in kidney and muscle
103 tissue, but the same concentrations in liver tissue (García et al., 2011). Higher Pb concentrations with
104 increasing age were found in male roe deer liver and muscle tissue in Spain (García et al., 2011),
105 while no age effect was found in bone or teeth in Poland. In red deer (*Cervus elaphus*) bone Pb
106 concentrations increase with age in Croatia (Lazarus et al., 2008), however, no such increase was
107 found in Spain (Rodríguez-Estival et al., 2013). Lazarus et al. (2018a) found no sex differences in Pb
108 concentrations in the femoral bones or in liver and kidney tissues (Lazarus et al., 2018b) of brown
109 bears (*Ursus arctos*) in Croatia, but animals ≥ 4 years had higher Pb concentrations compared to
110 younger individuals. Brown bear cubs-of-the-year (< 1 year old) had higher Pb concentrations in soft
111 tissue compared to yearlings, which indicates a transfer of Pb during pregnancy and lactation.

112 We used a brown bear population in Scandinavia as a sentinel for environmental Pb
113 exposure, and evaluated blood Pb concentrations in relation to life-history traits (age, body mass).
114 Studies on free-ranging brown bears in the USA, south-eastern Europe, and Scandinavia have

115 reported mean blood Pb concentrations of 55 $\mu\text{g/L}$ (Rogers et al., 2012), 61 $\mu\text{g/L}$ (Lazarus et al.,
116 2020), and 88 $\mu\text{g/L}$ (Boesen et al., 2019), respectively. These high Pb concentrations suggest that
117 brown bears may act as a good sentinel species. Brown bears are omnivorous and their diet typically
118 consists of vegetation and berries, but they also kill or scavenge on ungulates and feed on insects,
119 (Bojarska and Selva, 2012; Dahle et al., 1998; Rauset et al., 2012; Stenset et al., 2016; Swenson et al.,
120 2007b), and, thus, represent the cumulative burden of different potential Pb sources. Scandinavian
121 brown bears hibernate up to 6 months between October and April (Evans et al., 2016). Females mate
122 in June and delay implantation until the onset of hibernation in fall, and give birth in the winter
123 (Friebe et al., 2014; Tsubota et al., 1987). Females exhibit active-state body temperatures during
124 gestation, give birth after 56 days and then decrease their metabolic rate back to hibernation levels
125 (Friebe et al., 2014). In Sweden, offspring remain with their mothers for 1-2 years (Van de Walle et
126 al., 2018).

127 The aims of this study were to screen blood Pb concentrations in the brown bear population
128 in south-central Scandinavia and to evaluate if Pb concentrations are correlated with life-history
129 traits and lactation. We predicted that blood Pb concentrations in female brown bears increase with
130 age, body mass, and during lactation. We also tested for spatial correlations, because exposure to Pb
131 may vary in space and spatial clustering may affect life history traits differently. We further
132 hypothesized that brown bear offspring are exposed to Pb from lactation, and predicted that
133 variations in milk Pb concentration are related to a female's blood Pb concentration, and that the
134 offspring's blood Pb concentration is positively correlated with the mother's blood Pb concentration.

135 **Methods**

136 ***Study area***

137 This study is part of a long-term individual-based research project on brown bears in south-
138 central Sweden and south-eastern Norway ($\sim 61^\circ\text{N}$, 15°E) (Scandinavian Brown Bear Research Project,
139 2020). The size of the study area is approximately 13,000 km^2 , predominantly covered with
140 intensively managed coniferous forests in stands of different ages, ranging from recent clear cuts to
141 90-100-year-old stands (Martin et al., 2010; Swenson et al., 1999). The rolling landscape is
142 interspersed with lakes and bogs, and with agricultural fields towards the east. The altitude gradually
143 increases from ≈ 150 m above sea level in the east to 850 m above sea level in the west, which is also
144 the approximate tree line. Human settlements are concentrated in the north and south, with only
145 few high-traffic roads ($0.14\text{km}/\text{km}^2$). However, isolated houses (mainly cabins) and both paved and
146 gravel roads with low traffic volume are distributed throughout the study area (0.3 cabins/ km^2 and
147 0.7 km low-traffic roads/ km^2) (Martin et al., 2010).

148 Capture and sampling

149 All brown bear captures and sampling were carried out according to an established protocol
150 (Arnemo and Evans, 2017) approved by the Swedish Ethical Committee on Animal Research (Uppsala,
151 Sweden; Dnr 5.8.18-03376/2020), the Swedish Environmental Protection Agency (NV-00741-18), the
152 Swedish Board of Agriculture (#31-11102/12), the Norwegian Food Safety Authority (FOTS ID 19368)
153 and the Norwegian Environment Agency (2018/3346). All bears were darted from a helicopter in the
154 spring (April-May; 2010-2019), sex-determined and weighed using a digital spring scale. Because the
155 captures mainly focused on known females and their dependent yearling offspring, the age of most
156 captured individuals was known. Bears captured the first time as adults were aged by counting the
157 cementum layers of a vestigial first premolar (Mattson, 1993). All captured bears were tattooed and
158 microchipped for individual recognition.

159 Blood and milk sampling

160 Blood was collected from the jugular vein in 4 mL evacuated K3EDTA tubes (EDTA, n=118)
161 (Vacuette, Greiner Bio-One International GmbH, Kremsmünster, Austria) and in 6 mL evacuated
162 heparin trace element tubes (TE, n=54) (Vacuette). Mammary glands of adult females were palpated
163 to visually confirm lactation. We administrated 10 IU oxytocin (Vetocin 10 IU/mL, Bela-Pharm GmbH
164 & Co. Kg, Vechtra, Germany) to lactating females and collected approximately 1 mL milk in a 10 mL
165 non-collared screw cap tube (Sarstedt, Nümbrecht, Germany). Tubes with samples were frozen the
166 same day and kept at -20° C during storage and shipment to the laboratory (ALS Scandinavia AB,
167 Luleå, Sweden).

168 Pb analysis

169 At the laboratory, blood and milk samples were prepared for analysis by closed vessel
170 MicroWave-assisted acid digestion. Pb concentration in digests was measured by high-resolution
171 inductively-coupled plasma sector field mass spectrometry (ICP-SFMS, ELEMENT XR,
172 ThermoScientific, Bremen, Germany) using a combination of internal standardization and external
173 calibration. Quality assurance and quality control (QA/QC) included a set of preparation blanks and
174 matrix-matched control specimens (Seronorm Trace Elements Whole Blood Levels 1 and 2 from SERO
175 AS, Norway) prepared and analyzed with each analytical batch of blood samples. Contribution from
176 preparation blanks was less than 0.2 µg/L and thus negligible for Pb concentrations found in milk and
177 blood samples. Differences between found and target Pb concentrations for the controls were under
178 6%, the relative standard deviation (RSD) and of the same magnitude as typical instrumental
179 precision (in the range 3%-5% RSD). Further details on analytical methods can be found in Rodushkin
180 et al., 2000.

181 Sampled brown bears

182 We analyzed a total of 172 blood samples for Pb concentrations. The samples were collected on 153
183 sample events (2010: 13, 2013: 9, 2017: 31, 2018: 46, 2019: 37, and 2020: 17). Nineteen samples
184 were analyzed in pairs (collected at the same sample event). Sampled animals were comprised of
185 bears sampled during family group captures, i.e. mothers (N = 28) captured with 1-3 dependent
186 offspring (N = 56), family group capture attempts with either only the mother (N = 16) or only
187 offspring (N = 11) captured, and captures of single bears (N = 42). A total of 67 dependent offspring
188 were sampled, 55 yearling olds and 12 two-year olds. Mean age for both independent females and
189 males was 8.7 years (range: 3-25 years). We collected milk samples from 20 females; nine of those
190 had cubs of the year, eight had yearlings and three two-year old offspring. Individual bears were
191 sampled up to four times during the study period, a total of 110 individual bears are included in the
192 study.

193 Statistical analysis

194 To investigate possible contamination by the blood sampling tubes, we tested Pb
195 concentrations in 19 brown bears using both TE and EDTA tubes. We tested whether the differences
196 between Pb concentrations of the TE compared to the EDTA values were lower or greater than 0 with
197 a paired Wilcoxon rank test. We fitted a linear regression model with the TE value as the response
198 and the EDTA value as the predictor value and forced the intercept through zero. Pb concentrations of
199 blood collected in EDTA tubes were significantly lower than from samples collected in TE tubes (W =
200 151; P = 0.01), with a median difference of 2.51 $\mu\text{g/l}$ (SD= 2.46; range: 0.03 - 9.28 $\mu\text{g/L}$). We then
201 used the regression coefficient $\beta = 1.013$ to correct all EDTA based Pb concentrations in the data set;
202 after this correction, the newly calculated EDTA values were not significantly different from TE values
203 (W = 113; P = 0.49) and were used for further analysis. We evaluated the correlation between milk
204 Pb concentration and the blood Pb concentration with Spearman's rho. We further tested if milk Pb
205 concentration was related to the lactation period in years (i.e., the offspring's age) with a Kruskal-
206 Wallis test. We used R version 3.6.3 for all data handling and analyses (R Core Team, 2019).

207 Blood Pb variation in relation to life history and spatial correlation in female bears

208 To investigate spatial correlation of Pb values in relation to life history traits, we fitted a
209 variogram model with age and reproductive status as explanatory variables, using Pb concentration
210 as the identifier and 200 km as the cutoff value (i.e., the distance from the eastern to the western
211 edge of the study area). A second variogram was fitted and the result split by the four cardinal
212 directions. Variograms were fitted using the variogram function of the *gstat* library (Pebesma, 2018).

213 To test if spatially defined groups of sampled bears within the study area differed in their Pb
214 concentrations, we performed a hierarchical cluster analysis of the Euclidian distance between
215 capture locations using the *hclust* function and the complete method to find similar clusters from the
216 *stats* library in R. We plotted the cluster dendrogram and visually determined a reasonable cutoff
217 value (Zuur et al., 2009). We added the cluster ID to each observation in the data set.

218 We used a generalized linear model (GLM) to investigate whether blood Pb concentration of
219 independent (i.e. weaned from their mother) sub adult and adult females is affected by life history
220 traits, i.e. age (in years), reproductive status (lactating vs non-lactating), and the spatial cluster ID.
221 Because age and body mass are highly correlated in brown bears (Bartareau et al., 2011; Swenson et
222 al., 2007a; Zedrosser et al., 2006), we only used age as the explanatory variable. We fitted the GLM
223 with a gamma distribution and an identity link function to avoid estimation of negative Pb
224 concentrations. Despite repeated measurements, we decided not to include the bear ID as a random
225 variable due to non-convergence issues caused by too few replicates per bear ID (one to three
226 measurements per bear). The cluster ID variable was retained in all candidate models, except the
227 Null model. Then we fitted a set of candidate models comparing all possible variable combinations as
228 well as a Null model and carried out model selection based on the Akaike Information Criterion
229 corrected for small sample size (AICc). We averaged models within a cut off value of $\Delta AICc \leq 2$ with
230 the *AICcmodavg* package (Mazerolle, 2019). The mean gamma dispersion parameter from the two
231 models was used for model averaging. The age was known for only nine of the 15 solitary males and
232 we therefore decided to exclude males from this analysis.

233 ***Blood Pb correlation between mothers and offspring***

234 We used generalized linear mixed models (GLMM) with a Gaussian distribution and
235 maximum likelihood estimation (ML) to evaluate the blood Pb concentration of dependent offspring
236 in relation to the blood Pb concentration of their mothers. We used the offspring's Pb concentration
237 as the response variable, and the mother's Pb concentration as well as categories for sex and age of
238 the offspring as the explanatory variables. Because several offspring were captured as part of a
239 family group, we added the ID of the mother as a random intercept. We used a model with the blood
240 Pb concentration of the mother as a fixed variable as the base model and compared it to models
241 containing different combinations of the variables age and sex. We performed model selection based
242 on the lowest AICc value and averaged models within a $\Delta AICc \leq 2$ (Mazerolle, 2019).

243 **Results**

244 ***Sampling***

245

246 All 153 blood samples contained measurable concentrations of Pb. The overall mean blood
247 Pb concentration was $96.6 \mu\text{g/L} \pm 35.6 \mu\text{g/L}$ (SD) (Table 1), with a range from $38.7 \mu\text{g/L}$ measured in a
248 two-year-old male to $220.5 \mu\text{g/L}$ in an adult female.

249 We excluded two milk samples due to concerns by the laboratory over the validity of the
250 values because of very low collected volume and associated contamination risk. The Pb
251 concentration of the remaining milk samples ($N = 18$) ranged from $21.4 \mu\text{g/L}$ to $103.6 \mu\text{g/L}$, with a
252 mean of $42.9 \mu\text{g/L} \pm 21.1 \mu\text{g/L}$. Females with higher blood Pb concentration also had significantly
253 higher Pb concentration in their milk ($N = 18$, $\rho=0.60$, $p = 0.011$). The milk Pb concentration did not
254 differ between lactating females accompanied by cubs-of-the-year, yearlings, or two-year-old
255 offspring ($\chi^2 = 3.2$, $df = 2$, $p = 0.20$).

256 ***Blood Pb variation in relation to life history of female bears***

257 Age and reproductive status were available for 56 blood samples from 34 independent sub
258 adult and adult females (1-3 samples/individual). Thirty-one bears (age: 5-25 years) were lactating,
259 and 25 individuals (age: 2-10 years) were non-lactating. The plotted output of the variograms
260 displayed an approximately horizontal line over the entire distance tested and a lack of spatial
261 correlation (Figure S1). Splitting the variogram into the four cardinal directions revealed no clear
262 spatial patterns in any direction (Figure S2).

263 Based on the dendrogram of the cluster analysis, we chose a cut off value of 60 km resulting
264 in four different clusters named after their geographic location (from west to east: Fulufjell, Älvdalen,
265 Noppikosiki (Noppi) West and Noppi East; Figure 1 and Figure 2). Two models fell within $\Delta\text{AICc} \leq 2$;
266 the model with the lowest AICc contained reproductive status, age and the cluster ID as fixed terms,
267 and the second-best model contained the variables reproductive status and cluster ID as fixed terms
268 (Table 2). We used the model-averaged estimates of these two models for further interpretation
269 (Table 3). Compared to non-lactating females, the top model predicted a 1.3 to 1.6 times higher
270 blood Pb concentration in a lactating 10-year-old female bear, depending on cluster (Figure 3). Blood
271 Pb concentrations were highest in the cluster west from the study area center and decreased
272 towards the eastern cluster. Pb blood concentrations were also lower in older individuals, however,
273 with overlapping confidence intervals (CI's) and small effect size; e.g. the predicted Pb blood
274 concentration of a 15-year-old lactating female in the eastern-most cluster was $83.7 \mu\text{g/L}$ (95% CI:
275 $60.3\text{-}107.0 \mu\text{g/L}$) compared to $92.0 \mu\text{g/L}$ (95% CI: $76.5\text{-}107.6 \mu\text{g/L}$) for 10-year-olds (Table 3).

276 ***Blood Pb correlation between mothers and offspring***

277 Two GLMMs explaining blood Pb concentrations in offspring were within $\Delta AICc \leq 2$: the most
278 supported model contained only the maternal blood Pb concentration as explanatory variable, the
279 second model contained the offspring's age in addition. Further interpretation is based on the
280 averaged model output of these two models. Blood Pb concentrations of offspring increased
281 significantly (0.77 $\mu\text{g/L}$; 95% CI: 0.6-1.0 $\mu\text{g/L}$) with increasing maternal blood Pb concentration (Figure
282 4). For example, the model estimated a blood Pb concentration of 70.4 $\mu\text{g/L}$ (95% CI = 54.2-86.6
283 $\mu\text{g/L}$) for a yearling offspring if the mother has a blood Pb concentration of 82.0 $\mu\text{g/L}$ (1st quantile of
284 maternal Pb concentrations). If a mother has a blood Pb concentration of 138 $\mu\text{g/L}$ (3rd quantile of
285 maternal Pb concentrations), her yearling offspring has an estimated blood Pb concentration of 113.6
286 $\mu\text{g/L}$ (95% CI = 86.1-140.6 $\mu\text{g/L}$).

287

288 Discussion

289 We found that lactating females had significantly higher blood Pb concentrations compared
290 to non-lactating females (supporting prediction *i*). We also found support for the hypothesis that
291 offspring are exposed to Pb due to suckling, i.e. maternal Pb blood concentrations are significantly
292 and positively correlated with milk Pb concentration (supporting prediction *ii*) as well as with the
293 offspring's blood Pb concentrations (supporting prediction *iii*). The mean blood Pb concentration
294 (92.3 $\mu\text{g/L}$) of dependent offspring with a developing neurosystem exceeded the EFSA thresholds of
295 12 $\mu\text{g/L}$ for developmental neurotoxicity in children by a factor of eight, and all analyzed blood Pb
296 concentrations were above the threshold for increased systolic blood pressure of 36 $\mu\text{g/L}$ in humans
297 (EFSA, 2013). The blood Pb concentrations in Scandinavian brown bears (mean 96.6 $\mu\text{g/L}$) were also
298 higher than in brown bear populations of the greater Yellowstone area (mean 55 $\mu\text{g/L}$) and south-
299 eastern Europe (mean 61 $\mu\text{g/L}$) (Lazarus et al., 2020; Rogers et al., 2012). Our findings strongly
300 indicate high Pb exposure in the south-central Scandinavian brown bear population. Most likely,
301 bears are exposed to Pb throughout life, starting during the fetal period and continuing for their
302 entire life-spans at levels, considered a health risk in humans.

303 Lactation increases calcium turnover, and Pb stored in bones can be reabsorbed to the blood
304 and become an endogenous source for blood Pb (Silbergeld, 1991). In experimental studies on
305 laboratory mice, Keller and Doherty (1980) showed that the milk to plasma ratio is similar for calcium
306 and Pb. In humans, the relationship between maternal blood Pb concentration and milk Pb
307 concentration follows a linear relationship (Ettinger et al., 2014; Gulson et al., 1998). Mice with
308 intravenously injected Pb, transferred approximately 25% of the injected Pb dose to the suckling
309 pups (Keller and Doherty, 1980). This suggests that also lactating brown bears mobilize Pb from their

310 bones, resulting in increased blood Pb concentrations and in transfer to milk and excretion of Pb
311 from their bodies. A confounding factor in our study is that age and lactation are not independent.
312 Almost all females above 10 years of age were lactating at the time of sampling whereas only three
313 out of 23 females younger than 7 years old were lactating. Data exploration suggested an increasing
314 blood Pb concentration with increasing age (Figure 3). However, once we accounted for lactation, the
315 effect of age disappeared, indicating no correlation between blood Pb concentration and increasing
316 age.

317 More than one third of the offspring in the study area wean from their mothers during their
318 third year of life (Van de Walle et al., 2018) and we found that females with 2-year-old offspring were
319 still lactating (N=3). Although Pb concentration in milk decreases during the lactation period in other
320 mammals (Antunovic et al., 2005; Ettinger et al., 2014), female brown bears with 2-year-olds had
321 milk with similar Pb concentrations as females with cubs-of-the-year. We conclude that offspring are
322 exposed to Pb in the milk for the entire lactation period, i.e. up to 28 in our study population. Lazarus
323 et al. (2018b) found that renal Pb concentrations were higher in cubs-of-the-year compared to
324 yearling bears, which suggests a high initial Pb absorption capacity of younger offspring that
325 decreases with increasing age. We lack data from cubs-of-the-year, but similar to Lazarus et al.
326 (2018b), we found a high correlation between the mother's and the offspring's Pb concentrations. In
327 addition to Pb ingestion via milk, both the mother and her dependent offspring feed together on the
328 same resources and at the same locations for up to two years and have thus a similar environmental
329 Pb exposure.

330 Laboratory experiments on rats showed neurobehavioral effects in offspring of mothers with blood
331 Pb concentrations of 110 µg/L (Virgolini et al., 2008). The blood Pb concentrations of bear offspring
332 in our study ranged from 38.7-166.5 µg/L, i.e. Pb concentrations related to behavioural changes in
333 other mammalian species (Ma, 2011).

334 Blood Pb concentrations have been used as an indicator of recent Pb exposure due to an
335 estimated half-life of 35 days in humans (Rabinowitz et al., 1976). Due to the reabsorption from
336 bones and tissue and the equilibrium between various body compartments, however, the actual
337 blood Pb concentration might take much longer to halve. In domestic cattle exposed to a high initial
338 dose of Pb, blood concentrations halved after 68 to 266 days, and in chronically exposed humans
339 after a median of 619 days (Hryhorczuk et al., 1985; Miranda et al., 2006). Increased blood Pb
340 concentrations measured at the population level, as in brown bears in Scandinavia, Eastern Europe
341 and Yellowstone, are most likely due to chronic Pb exposure.

342 Much of the Pb in the environment of the Scandinavian brown bears is from aerial
343 depositions originating in emissions from leaded gasoline and smelters from entire Europe (Renberg
344 et al., 2001; von Storch et al., 2003). A possible exposure pathway are major food resources, such as
345 bilberries and lingonberries that take up Pb from the soil. According to Welch et al. (1997), a bear of
346 80 kg body mass needs 0.925 kg wet weight daily intake of bilberries to maintain body mass and has
347 a maximum digestive capacity of 28.6 kg wet weight per day. Rodushkin et al. (1999) reported
348 background Pb concentration of 3.4 µg/kg wet weight in bilberries. For a brown bear of 80 kg the
349 daily Pb ingestion from berries would be 0.04 µg/kg to maintain body mass and 1.22 µg/kg body
350 mass for a bear consuming berries at the digestive capacity.

351 Experimental studies evaluating the amount of ingested Pb in relation to blood Pb
352 concentration focus on exposure, i.e. subjects are exposed to contaminated food, water or air/dust
353 with the total ingested dose unknown. For example, in laboratory rats, a 2-month intake of drinking
354 water with a Pb concentration of 50 µg/L resulted in a blood Pb concentration of about 110 µg/L
355 (Cory-Slechta et al., 1983). In comparison, the mean Pb concentration in bear milk in our study is 44.9
356 µg/L and the mean blood concentration of dependent offspring is 93.2 µg/L, suggesting similar
357 exposure compared to the laboratory rats from Cory-Slechta et al., 1983. In laboratory mice, a 3-
358 month exposure to food with 200 µg Pb/kg dry mass led to a blood Pb concentration of 70 µg/L
359 (Iavicoli et al., 2003). In a pilot study (Fuchs, unpublished), we sampled berries inside the most
360 eastern cluster (Figure 1), and found mean dry weight concentrations of 9.0, 6.7 and 30.8 µg/kg for
361 bilberries, lingonberries and crowberries, respectively. The mean blood Pb concentration of 96.6 µg/L
362 in bears is similar to concentrations measured in laboratory rodents, however, exposure from berries
363 is likely much lower.

364 Rabinowitz et al. (1976) exposed five humans to a daily dietary Pb ingestion ranging from 2.4
365 µg to 5.1 µg Pb/kg body mass and were able to maintain each subject's pre-study Pb concentrations
366 ranging between 170 µg/L and 250 µg/L blood. Reported ingestion in Rabinowitz's study was 2 to 4
367 times higher than Pb ingestion of a bear with a berry intake at the digestive capacity, similar to the
368 human subject's blood Pb concentrations which were 2 to 3 times higher compared to the mean
369 blood concentration of bears in this study. However, this assumes similar uptake rates and Pb
370 kinetics between humans and bears. In addition, while Rabinowitz et al. (1976) sampled during
371 ingestion, the bears were sampled after hibernation and 7 to 8 months after the peak hyperphagic
372 berry consumption.

373 Bears that consume vegetation and insects, such as ants, commonly ingest significant
374 amounts of soil, humus and other debris (Stenseth et al., 2016). Bears might hide larger prey with

375 vegetation and soil, probably ingesting residues when resume feeding. Salminen et al. (2005)
376 reported increased Pb concentrations in humus in relation to sub and top-soils due to aerial
377 depositions for our entire study area. Further, the increased blood Pb concentration predicted west
378 in the study area for female bears coincides with increased Pb concentrations in sub soil in the same
379 area (Salminen et al., 2005). It remains unclear if and how, hibernation affects Pb kinetics and blood
380 Pb concentration in bears. During hibernation, there is no ingestion of Pb and excretion of Pb is likely
381 very low because bears do not eat or drink during hibernation, their urine production is very low and
382 bile excretion entirely absent (Nelson et al., 1973).

383 Large game hunting with Pb-based ammunition poses an additional potential source for high
384 blood and milk Pb concentrations of bears. In Sweden, 90% of bullets used to harvest moose are Pb
385 based and these bullets normally fragment into inner organs that are left for scavengers (Stokke et
386 al., 2017). The moose-hunting season in central Sweden starts in early September and lasts until the
387 end of January. While most bears start hibernating in late October (Evans et al., 2016), pregnant
388 females may enter their dens between late September and late October (Friebe et al., 2014, 2001;
389 Manchi and Swenson, 2005). Of the total moose harvest in our study area, 75% are shot during
390 October and consequently the available moose-based biomass for scavengers is estimated to be 15
391 fold higher compared to before the hunting season (Länsstyrelserna, 2019; Wikenros et al., 2013).
392 We lack data on the frequency of use of gut piles by brown bears but individuals might have reduced
393 mobility or hibernate through parts of the peak hunting season. During spring in Scandinavia, after
394 hibernation, moose is a major proportion of the bears estimated dietary energy content (Stenset et
395 al., 2016). Winter mortality, traffic kills and, in areas with sympatric wolf (*Canis lupus*) presence, wolf-
396 killed moose compose the available biomass in spring (Ordiz et al., 2020; Wikenros et al., 2013). A
397 potential source of Pb exposure during spring are slaughter remains that are discarded in the forest
398 by hunters and visited by bears. Common ravens (*Corvus corax*) in Wyoming USA, showed sharp
399 increases in blood Pb concentrations by the start of the moose hunt and slightly elevated
400 concentrations during the snow melt in spring when hunting remains from the fall reappear
401 (Craighead and Bedrosian, 2008). Gut piles from the fall however, are commonly rapidly consumed
402 and only rarely available to bears in the spring (Gomo et al., 2017).

403

404 **Conclusions**

405 Scandinavian brown bears are highly exposed to environmental Pb despite the generally large
406 decrease of the Pb burden on a spatio-temporal scale in Europe. In Scandinavia, sediment core Pb
407 concentrations dated back to pre-human metallurgic activities are very low (Renberg et al., 2001)

408 therefore we assume that the major proportion of Pb exposure in bears is related to human use of
409 Pb. The generally high Pb blood concentrations of bears and the differences in Pb concentrations
410 between lactating and non-lactating females indicate endogenous release of Pb stored in other body
411 parts into the blood. The exogenous sources remain unclear, but berry consumption, ingestion of soil
412 during foraging and Pb from large game hunting with Pb-based ammunition are most likely the major
413 contributing factors. For suckling offspring, Pb in milk is most likely the major source. Based on the
414 Pb blood concentrations observed the early life stages and latent exposure on the population level,
415 we would expect to observe adverse effects at individual or population level. Our findings are in
416 contradiction to the dramatic decrease in Pb exposure in Scandinavia over the past decades in both
417 the environment and humans as a result of successful mitigation of Pb pollution from gasoline
418 additives. We see a strong need to investigate the source of the Pb exposure in the bears and to
419 investigate potential negative health effects in order to further reduce Pb pollution in the
420 environment.

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427 **Author contributions**

428 **Arnemo:** Conceptualization; **Fuchs, Thiel:** Methodology; **Rodushkin:** Resources; **Arnemo, Boesen,**
429 **Evans, Fuchs, Græsli, Hydeskov, Rodushkin, Thiel:** Investigation; **Fuchs, Thiel:** Formal analysis; **Fuchs:**
430 Writing – Original Draft; **Arnemo, Brown, Evans, Fuchs, Græsli, Hydeskov, Kindberg, Rodushkin,**
431 **Thiel, Zedrosser:** Writing – Review & Editing; **Arnemo, Zedrosser:** Supervision; **Arnemo, Kindberg:**
432 Project administration; **Arnemo, Kindberg:** Funding acquisition

433 **Competing interests**

434 I. Rodushkin is employed by ALS Global AB in Luleå, Sweden.

435 **Literature**

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Table 1: Mean blood lead (Pb) concentrations ($\mu\text{g/L}$) of brown bears (*Ursus arctos*) in Scandinavia, collected in 2010, 2013, and 2017-2020. Concentrations are shown for all bears in the data set (All), solitary males and females, females accompanied by dependent offspring, and dependent offspring. N = sample size, mean = arithmetic mean, SD = standard deviation of the mean, range = minimum and maximum values observed in the data.

Group	N	Mean	SD	Range
All	153	96.6	35.6	38.7 – 220.5
Solitary	42	87.2	32.9	40.9 – 175.2
- males	15	104.2	42.4	41.9 – 175.2
- females	27	77.7	21.9	40.9 – 132.3
Females with dependent offspring	44	112.0	37.5	53.1 – 220.5
All offspring	67	92.3	33.0	38.7 – 166.5

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Table 2: Model selection for estimating blood lead (Pb) concentration ($\mu\text{g/L}$) in female brown bears (*Ursus arctos*) in Scandinavia, based on the Akaike Information Criterion for small sample size (AICc). Estimated parameters are lactating or non-lactating bears (Status), the Age of the bear in years, and the ID of the spatial Cluster, the number of estimated parameters (K), the AICc, the AICc difference to the top model (ΔAICc) and the weight of evidence (W_i) for each model. Data was collected in 2010, 2013, and 2017-2019.

Model	Parameters	K	AICc	ΔAICc	W_i
Full	Status + Age + Cluster	7	518.63	0	0.71
Status	Status + Cluster	6	520.47	1.84	0.28
Age	Age + Cluster	6	535.86	17.23	0.00
Cluster	Cluster	5	536.25	17.62	0.00
Null	1	2	546.36	27.73	0.00

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Table 3: Generalized linear regression model estimates predicting blood lead (Pb) concentrations ($\mu\text{g/L}$) for female brown bears (*Ursus arctos*) in Scandinavia for the study period 2010, 2013, and 2017-2019. Status Lactating: Lactating female with offspring, Age: Age of the bear in years, Cluster: ID of the four spatial clusters based on spatial cluster analysis (Cluster Fulufjell is used as intercept).

Coefficient	Estimate	Standard Error	95% Confidence intervals
Intercept	100.00	11.89	76.69 – 123.32
Status Lactating	35.98	10.01	16.35 – 55.60
Age	-1.68	0.79	-3.23 – -0.12
Cluster Älvdalen	28.94	20.77	-11.76 – 69.64
Cluster Noppi West	-16.58	11.20	-38.54 – 5.37
Cluster Noppi East	-27.17	11.39	-49.50 – -4.84

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Table 4: Mixed linear regression model estimates predicting blood lead (Pb) concentrations ($\mu\text{g/L}$) for offspring brown bears (*Ursus arctos*) in Scandinavia for the study period 2010, 2013, and 2017-2020, dependent on the blood Pb concentration of the mother and whether offspring is one or two years old. The mothers ID was used as a random factor.

Coefficient	Estimate	Standard Error	95% confidence interval
Intercept	7.48	11.42	-15.29 – 30.24
Blood Pb Mother	0.77	0.10	0.57 – 0.97
Offspring age: 2 years	7.41	6.22	-4.79 – 19.60

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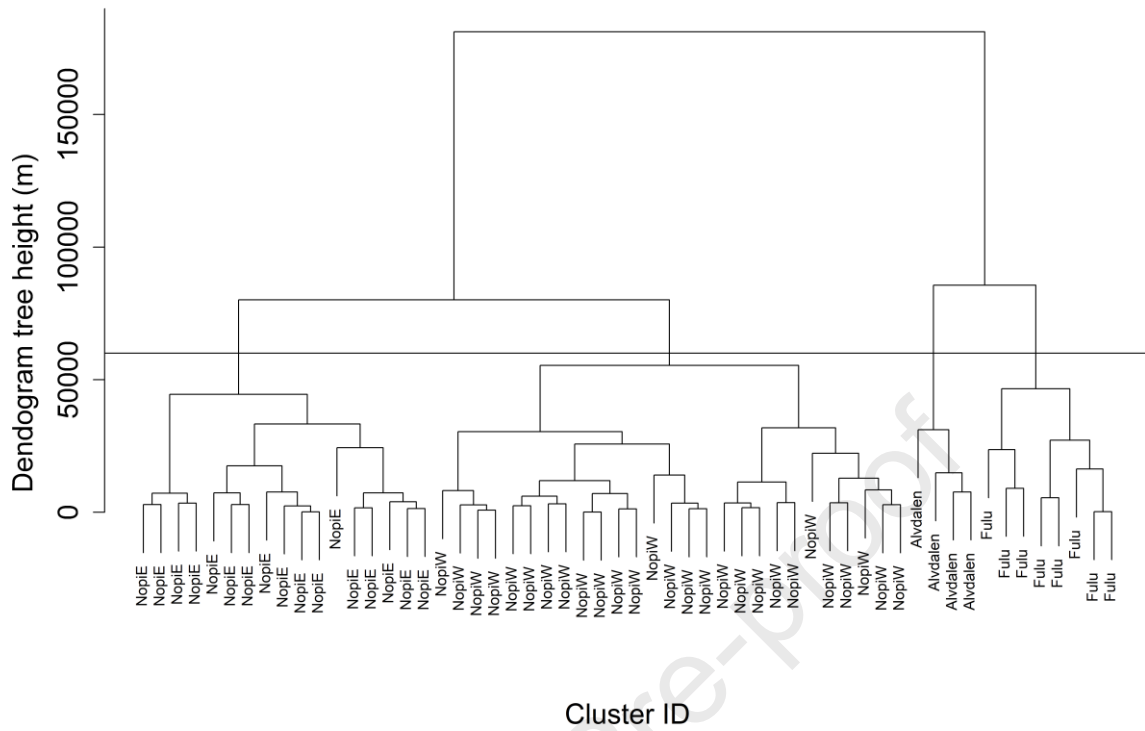


Figure 1: Dendrogram of the hierarchical cluster analysis based on brown bear (*Ursus arctos*) sampling locations from Scandinavia in 2010, 2013, and 2017-2019. Locations grouped below the cutoff value of 60,000 meters (solid horizontal line) build a spatial cluster. The cluster ID is added to the dataset and included in the linear model.

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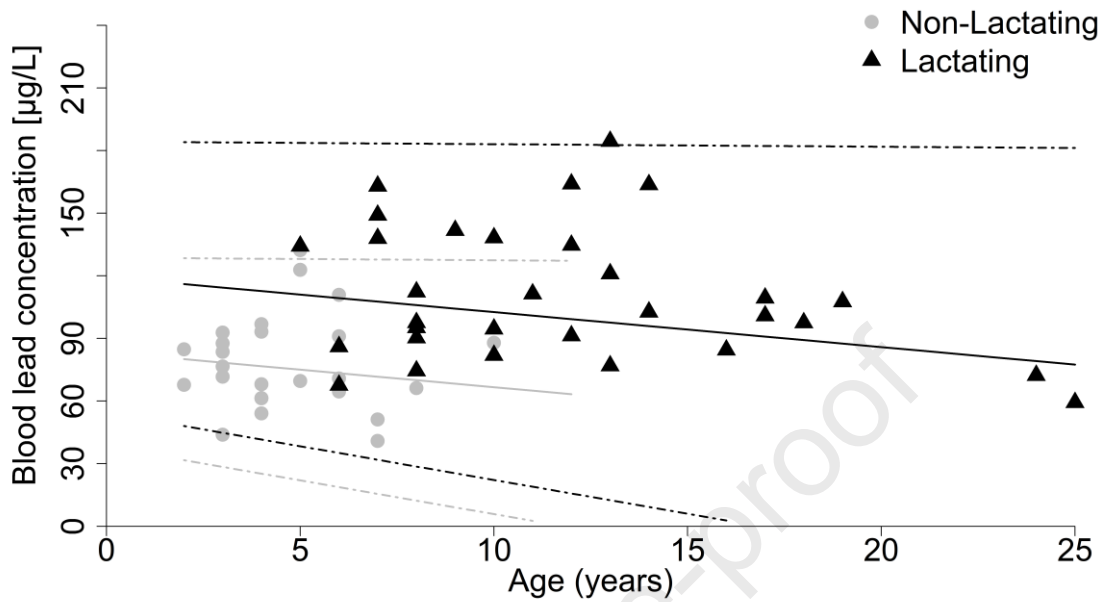


Figure 3: Blood lead (Pb) concentrations of lactating (black triangle) and non-lactating (grey dot) adult female brown bears (*Ursus arctos*) in Scandinavia in relation to their age. Model predictions for a lactating bear (black solid line) with 95% confidence interval (CI; black dash-dotted lines) and non-lactating bears (grey solid line, 95% CI grey dash-dotted lines) in the Noppi West cluster. Data collected in 2010, 2013, and 2017-2019.

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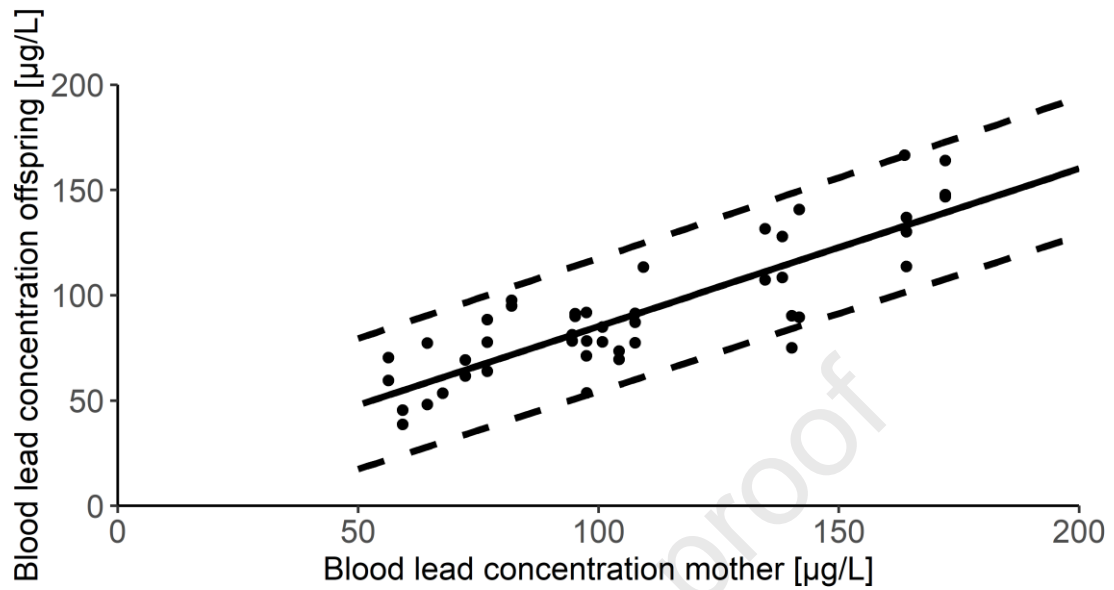


Figure 4: Blood lead (Pb) concentrations of Scandinavian brown bear (*Ursus arctos*) offspring in relation to the blood lead concentration of their mother (points) and predicted values for female offspring from a linear mixed model (solid line) with 95% confidence intervals (dashed lines). Data collected in 2010, 2013, and 2017-2020.

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Highlights

- Blood and milk lead (Pb) of Scandinavian brown bears was quantified by ICP-SFMS
- In lactating female bears, milk Pb was positively correlated with blood Pb
- Blood Pb of suckling cubs was positively correlated with their mothers blood Pb
- Dependent cubs are exposed to milk Pb throughout their suckling period
- Scandinavian brown bears are highly exposed to environmental Pb

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Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

I. Rodushkin is employed by ALS Global AB in Luleå, Sweden.

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